



Effects of Low-Level Artificial Light at Night on Kentucky Bluegrass and an Introduced Herbivore

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Increasing evidence suggests that artificial light at night (ALAN) can negatively impact organisms. However, most studies examine the impacts of ALAN on a single species or under high levels of artificial light that are infrequent or unrealistic in urban environments. We currently have little information on how low levels of artificial light emanating from urban skyglow affect plants and their interactions with herbivores. We examined how short-term, low levels of ALAN affect grass and insects, including growth rate, photosynthesis, and stomatal conductance in grass, and foraging behavior and survival in crickets. We compared growth and leaf-level gas exchange of Kentucky Bluegrass (*Poa pratensis*) under low-levels of ALAN (0.3 lux) and starlight conditions (0.001 lux). Furthermore, each light treatment was divided into treatments with and without house crickets (*Acheta domesticus*). Without crickets present, bluegrass grown under ALAN for three weeks grew taller than plants grown under natural night light levels. In the fourth week when crickets were introduced, grass height decreased resulting in no measurable effects of light treatment. There were no measurable differences in grass physiology among treatments. Our results indicate that low levels of light resulting from skyglow affect plant growth initially. However, with herbivory, the effects of ALAN on grass may be inconsequential. Gaining an understanding of how ALAN affects plant-insect interactions is critical to predicting the ecological and evolutionary consequences of anthropogenic light pollution.

Keywords: photosynthesis, urban light, growth rate, crickets (Gryllidae), photobiology

INTRODUCTION

Artificial light at night (ALAN) is an anthropogenic pollutant that is increasing spatially by a rate of 2.2% per year (Kyba et al., 2017). Direct ALAN sources, such as streetlights, can lead to skyglow: the atmospheric scattered light that can propagate up to several hundred kilometers into the environment (Luginbuhl et al., 2009; Aubé, 2015). Skyglow results in light encroaching into natural areas where direct sources of light pollution are not present (Gaston et al., 2015; Garrett et al., 2020).

The study of ALAN as an anthropogenic pollutant is a relatively young field (Longcore and Rich, 2004; Seymoure, 2018; Dominoni et al., 2020; Sanders et al., 2021), with most studies conducted at relatively high levels of nocturnal light pollution (e.g., 10–100 lux) (Gaston et al., 2013) but see Alaasam et al. (2018); Sanders and Gaston (2018). For reference, a full moon could create ambient light levels of 0.3 lux on its brightest nights (Biberman et al., 1966; Kyba et al., 2017). These high light levels are representative of organisms functioning under direct light pollution, such as directly beneath a streetlight, whereas most urban environments exist at lower light levels due to skyglow (e.g., 0.1–1.0 lux), which can impact environments several hundred kilometers away from a direct light source (Gaston et al., 2013; Dominoni et al., 2014; Seymoure et al., 2019a). Therefore, examining the impacts of light pollution at high intensities, although informative, is not representative of artificial light conditions in urban habitats at night. It remains an open question as to whether low levels of skyglow illumination (0.001–0.3 lux) affects communities to the same extent as direct illumination.

The intensity and spectral composition of light depends upon the phase of the moon, season, and weather, all of which create necessary cues for organisms (Kyba et al., 2015; Spitschan et al., 2016; Seymoure et al., 2019b). Plants use light as a cue for almost every physiological process including, but not limited to, seedling development, photosynthesis, growth, and budding (Briggs and Christie, 2002; Takemiya et al., 2005; Bennie et al., 2016; Gaston et al., 2017; Singhal et al., 2018). In addition to powering the electron transport chain in thylakoid membranes, light intensity and direction impacts photosynthetic efficiency through phototropism (i.e., the movement of the plant toward sunlight; Celaya and Liscum, 2005), chloroplast movements (Wada et al., 2003), and light-induced stomatal opening to optimize water-use efficiency (Dietrich et al., 2001). Periods of darkness are also important for plant metabolic processes, particularly stress recovery, including recovery from herbivory events (McNaughton, 1983; Singhal et al., 2018).

Increased levels of ALAN from urbanization are changing natural light regimes by increasing the intensity and duration of light available at night (Davies et al., 2013; Seymoure et al., 2019a; Buxton et al., 2020), potentially altering plant-herbivore interactions. For example, by masking natural night light levels, ALAN can mislead herbivores to be more active at night and disrupt plant-herbivore interactions and critical dark recovery periods for plants (Dominoni et al., 2020). Plants in light polluted environments may experience changes in pollination, photoreceptor signaling, phenology and flowering (Ffrench-Constant et al., 2016; Singhal et al., 2018), which can have ecological consequences for food web dynamics (Polis et al., 2004). However, little is known about how constant illumination at the level of urban light alters plant-insect interactions. ALAN has led to declines in population sizes of a diversity of insect species through its interference with insect development, movement, foraging, and reproductive success, which can alter trophic systems (Owens and Lewis, 2018; Owens et al., 2020).

Here we test whether short-term exposure to ALAN affects plant-insect interactions by modifying plant photobiology and

growth rates. We exposed two common urban species—Kentucky bluegrass (*Poa pratensis*), a cool season common turfgrass (Weissman and Rentz, 1977; Read et al., 1999; Suplick-Ploense and Qian, 2005), and the house cricket (*Acheta domesticus*), a nocturnal herbivore—to starlight (0.001 lux) and realistic urban nighttime light levels (0.3 lux) (Dominoni et al., 2013; Alaasam et al., 2018; Seymoure et al., 2019a) in order to test the following hypotheses: (1) Low levels of ALAN affect plant physiology. We predicted that plants grown under urban light would have higher net photosynthesis and dark respiration, increased growth rates, and increased stomatal conductance compared to control plants grown under starlight conditions. (2) Herbivory interacts with ALAN to affect plant biomass. We predicted cricket herbivores would reduce the biomass and height of grass. However, as crickets are nocturnal foragers, we predicted they would consume less plant material under urban light than starlight conditions and have lower survival rates in urban light.

MATERIALS AND METHODS

Light Treatments

We used a CMP6050 growth chamber (BDR16, Version 4.06, Conviron, Winnipeg, Manitoba) set to a temperature of 22.2°C with light control to create artificial light environments (0.3 lux, hereafter “urban light”) and natural new moon light environments (0.001 lux, hereafter “starlight”) at night (Dominoni et al., 2013; Alaasam et al., 2018; Seymoure et al., 2019a; Jones et al., 2020). Daytime light levels were 135, 300 lux, which is similar to natural daytime lux levels. There were two different light types in the chamber—high pressure sodium and mercury vapor—placed in alternating positions on the ceiling of the chamber. This is standard for the CMP6050 growth chamber. These lights were stepped up and down to simulate dawn and dusk in the chamber. Standard LED lights of the chamber remained off to create more realistic and desired light levels. To create urban light levels within the chamber, we used 4 layers of filter gels over the light sources (Rosco E-Colour #211.9 Neutral Density Filter, Stamford, CT, United States) that attenuated 83% of light. To further attenuate light, 90% black shade cloth was placed over starlight treatments, and 22% white shade cloth was placed over urban light environments. Shade cloth and filter gels only effect the quantity of light, but not the quality of light in the chamber. These were constructed as square boxes and placed over the plant treatment groups using PVC pipe and shade cloth. We confirmed that light levels were approximately 0.3 lux and 0.001 lux using a highly sensitive spectroradiometer (StellarNet Silver Nova, Tampa Bay, FL, United States) with a cosine corrected irradiance probe affixed to a 1000-micron optical fiber (StellarNet, Tampa Bay, FL, United States). We checked irradiance measurements using SpectraWhiz software (StellarNet, Tampa Bay, FL, United States); due to the low light levels, we set integration time to approximately 20 s for the 0.3 lux measurements and 8 min for the 0.001 lux measurements. This confirmed that light levels throughout the enclosure were within one order of magnitude of the chosen light level for each treatment: 0.3 and 0.001 lux.

Experimental Design

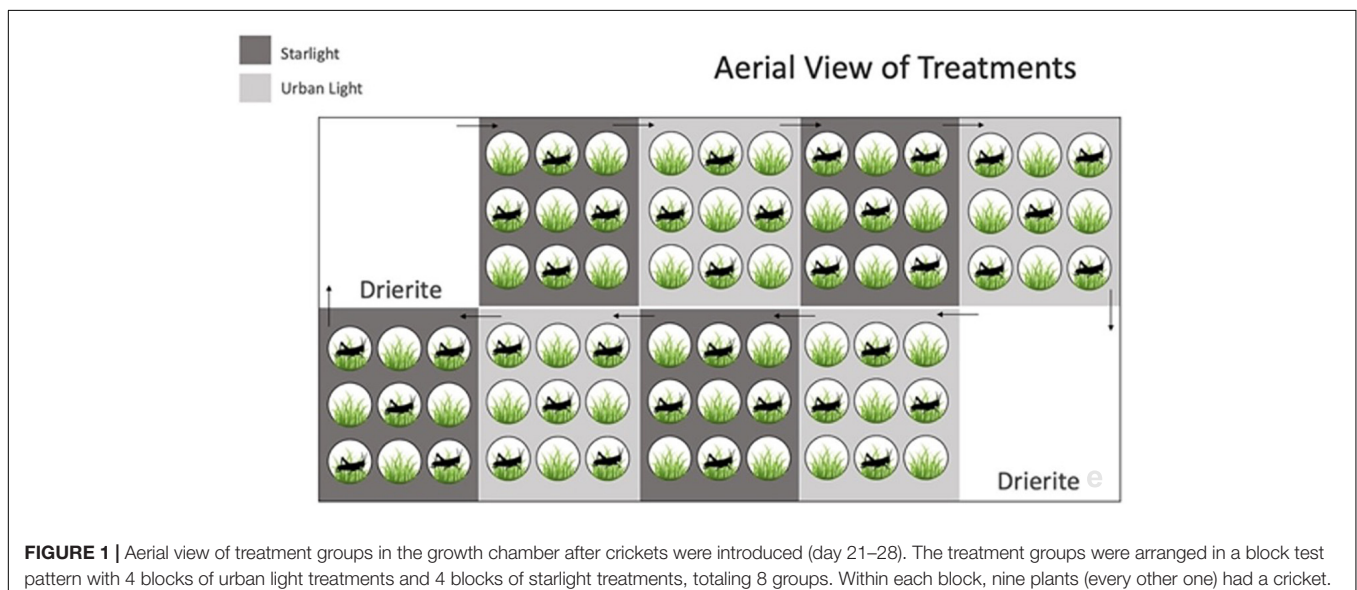
On day 1, Kentucky bluegrass seeds were sown in 10 cm round pots ($n = 72$) containing Scotts Miracle-Gro soil and placed in the growth chamber under experimental light conditions. On day 21, we measured the tallest blade of grass, then weeded down the pots randomly, excluding the tallest blade of grass, until there were 25 shoots of grass remaining. Weeding to a standard number of shoots ensured that there were no differences in grass abundance among measurements prior to the experiment (Lemoine et al., 2018). After the initial 21-day growth period, one randomly selected juvenile cricket, male or female, was placed in each of 36 designated cricket pots. Herbivory and light environments were examined using a 2×2 factorial design in which light treatment was factorially crossed with cricket treatment in a 28-day experiment. The four treatments were arranged in a block test pattern, as shown in **Figure 1**. Treatment groups included: (1) plants without crickets in urban light, (2) plants without crickets in starlight, (3) plants with crickets in urban light, and (4) plants with crickets in starlight ($n = 18$ per treatment). Nighttime lighting conditions were imposed in the middle of the day from the start of the experiment to ensure nighttime measurements could be taken during regular working hours. Lighting conditions were altered twice daily; we placed filter paper and shade cloth structures over the plants at 08:00 and removed them at 18:00 to create a 14:10 light:dark cycle typical of summer in the northern hemisphere. Blocks were rotated daily one position clockwise to account for spatial variation in light levels within the chamber, and generously watered at this time. Drierite (W.A. Hammond 23005, Xenia, OH, United States) was placed in two trays on opposite sides of the chamber to control humidity and prevent mold growth (Hammond, 1935).

Crickets were sourced as juveniles from a stock population from Premium Crickets (Winder, Georgia) in December 2018 as juveniles at a mean size of 1.9 cm, before the adult phase. From day 21 to 28, cricket survival was monitored daily (i.e.,

when light conditions were shifted) and categorized as alive or dead. All crickets were juveniles from day 21 to day 28 and thus we only report survival of juveniles. If a cricket was found dead, the cricket and its designated plant were removed from the experiment. Upon removal, we measured the height of the tallest blade of grass and recorded the length of time the plant/cricket spent in the chamber. We also cut and weighed aboveground biomass to determine wet and dry mass. On day 28, we removed all remaining plants from the experiment and recorded the final height of the tallest blade of grass. We calculated the average daily growth rate in week four (day 21 to day 28) to control for plants that were removed prematurely due to cricket death.

Gas Exchange Measurements

To assess the effect of light treatment on bluegrass physiology independent of herbivory, we measured leaf photosynthetic responses on day 19 before crickets were placed into pots. We measured leaf gas exchange in each light treatment using a LI-6400XT infrared gas analyzer with a leaf chamber fluorometer attached (Li-Cor Biosciences; Lincoln, NE, United States) following previously published methods with slight modifications (Lemoine et al., 2018). Plants were removed from the growth chamber temporarily for gas exchange measurements. The environmental conditions inside the leaf chamber were standardized across measurements; leaf temperature was maintained at 20°C, relative humidity was maintained between 40 and 50%, sample chamber flow rate was set to 200 $\mu\text{mol m}^2 \text{s}^{-1}$, and reference chamber CO_2 concentration was set to 400 ppm. Low flow settings are commonly used for small leaved grasses with low photosynthetic rates (Taylor, 2014). Leaf level gas exchange was measured under two light conditions: dark and low light (10 $\mu\text{mol m}^2 \text{s}^{-1}$ (740 lux) photosynthetically active radiation; PAR). Gas exchange in the dark provides an estimation of leaf respiration. The low light level was the minimum amount of light provided by the Li-6400 light source;



thus, we were unable to measure photosynthesis under the tested ALAN conditions imposed here ($<10 \mu\text{mol}$, $<740 \text{ lux}$), but instead measured whether treatments had an impact on plant photosynthetic responses to low levels of light. A newly emerged and fully expanded leaf from each individual ($n = 10$ individuals per treatment) was inserted into the leaf chamber. Prior to measurements, leaves were dark adapted for 2 h under a dark box that allowed no light to enter. Leaves were left in the chamber for 2–5 min to equilibrate to chamber conditions before gas exchange parameters (photosynthesis or respiration, and stomatal conductance) were recorded (average of three logged values taken in rapid succession). Steady-state fluorescence (F_s) was measured continuously before exposing plants to a saturating pulse of light ($2,750 \mu\text{mol m}^{-2} \text{ s}^{-1}$ of blue light or $\sim 203,500 \text{ lux}$) (Thimijan and Heins, 1983) to measure maximum chlorophyll fluorescence. Light inside the chamber was then switched to the low light level ($10 \mu\text{mol m}^{-2} \text{ s}^{-1}$). Once gas exchange reached stability, net photosynthetic rate and stomatal conductance were recorded, and a saturating pulse was applied to estimate photosystem II efficiency (ΦPSII): $\Phi\text{PSII} = (F_m' - F_s)/F_m'$ where F_m' represents chlorophyll fluorescence under low light. As grass blades rarely filled the entire chamber, the measured leaf area was estimated using width and length, and photosynthetic parameters, which are based on the area of the chamber (6 cm^2), were adjusted accordingly.

Data Analysis

All statistical analyses were performed in R version 3.4.3 (R Development Core Team, 1999). We first confirmed that our data were normally distributed to enable the use of parametric tests. To test our first hypothesis that gas exchange increased under ALAN, we ran a MANOVA with net photosynthetic rate, stomatal conductance, dark respiration, and ΦPSII as response variables and with light treatment and block as explanatory variables (Figure 2). For our second hypothesis that light and cricket treatments would affect plant height, we modeled daily percent change in height between day 21 and day 28 using a two-way ANOVA with light treatment, cricket treatment, and block as explanatory variables (Figure 3). We then analyzed the data using two-way ANOVA, again with light treatment, cricket treatment, and block as explanatory variables, testing for an interaction between light treatment and cricket treatment. We also analyzed cricket survival using Kaplan-Meier analysis with the “survival” package in R (Figure 4) (Therneau and Lumley, 2009).

RESULTS

There was no difference in net photosynthesis, stomatal conductance, dark respiration, or ΦPSII between grass grown in the two light treatments (Table 1). On day 21, bluegrass grown in urban light was taller (mean = 5.35, $\text{sd} = 1.02$) than bluegrass grown in starlight (mean = 4.79, $\text{sd} = 0.63$, Table 2). However, the daily percent change in plant height from day 21 to day 28 was not significantly different between treatments (Table 3). The presence

of crickets did affect plant height, whereby bluegrass with crickets present were shorter than bluegrass without crickets (Table 3).

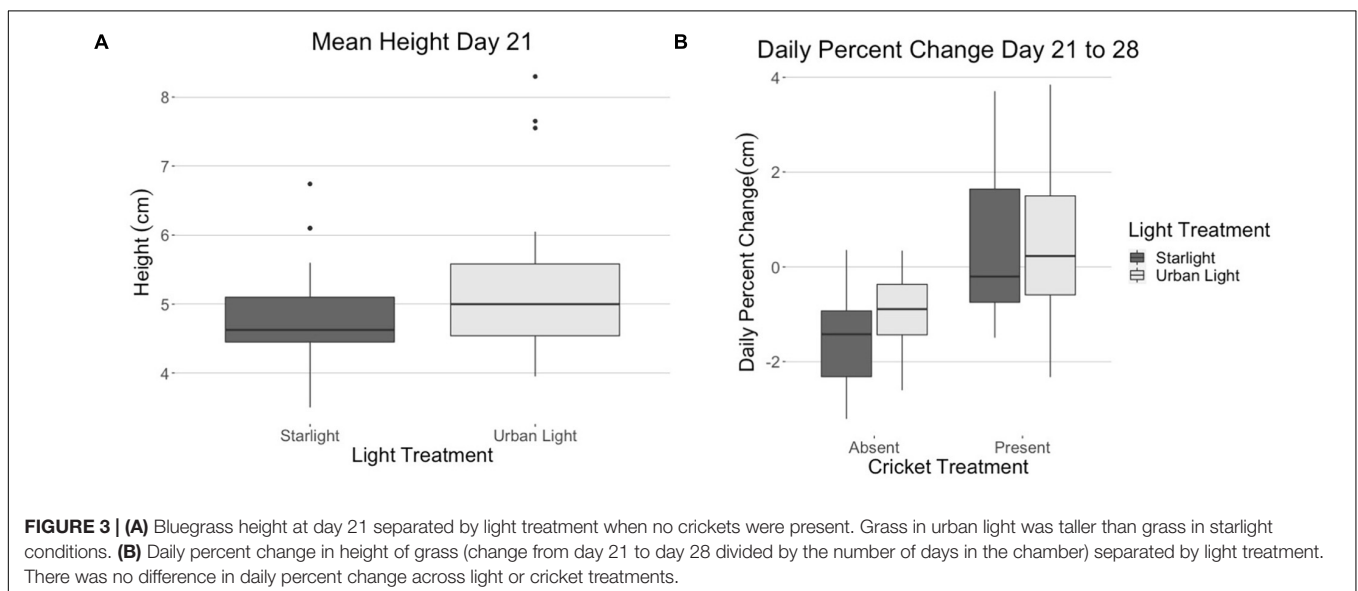
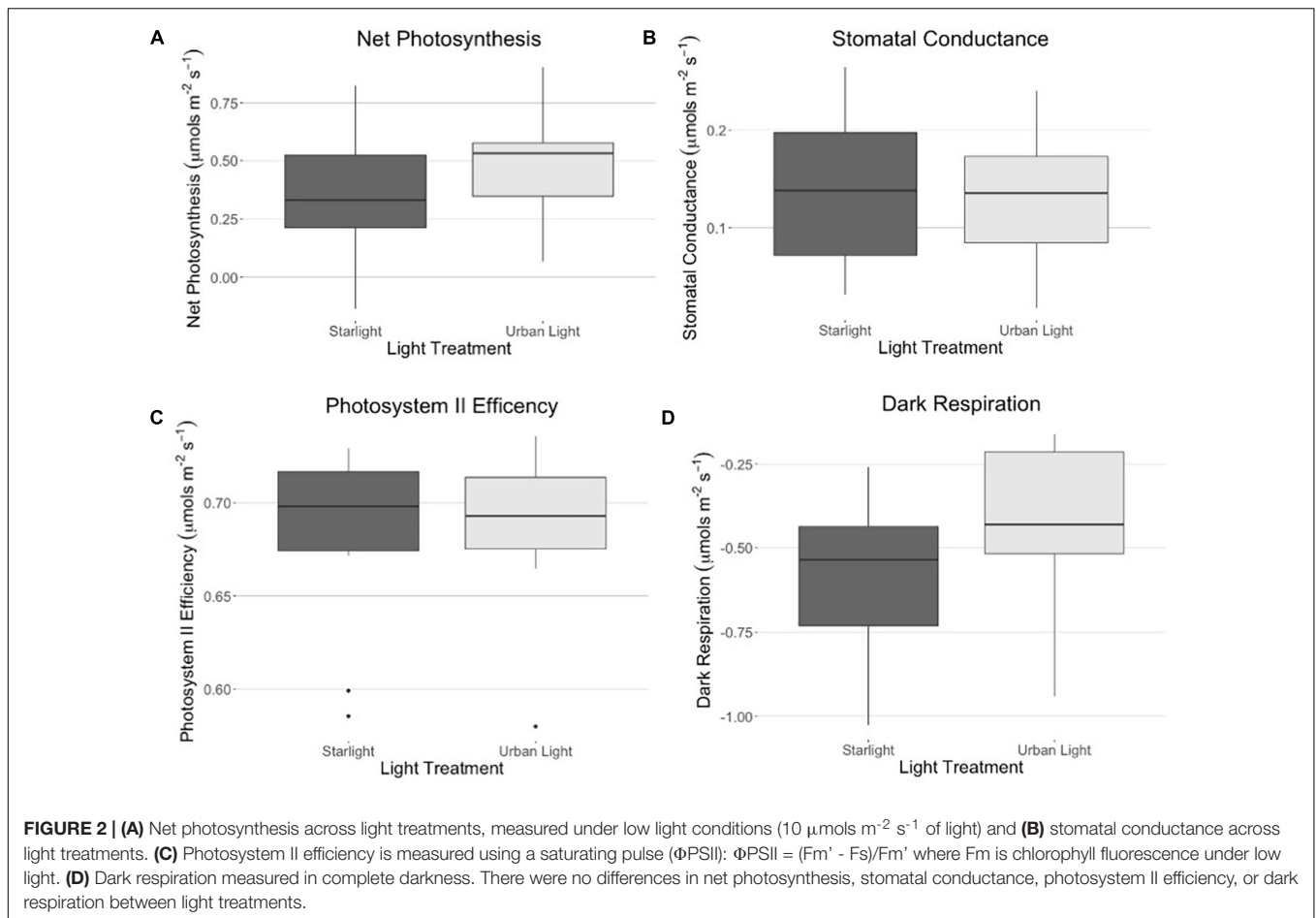
Crickets in the urban light treatment had a 25.0% probability of survival, whereas crickets in the starlight treatment had a survival probability of 32.1%, but this difference was not statistically significant (Kaplan-Meier: $n = 36$, $p = 0.37$, Figure 4). There was no difference in survival due to sex (Kaplan-Meier: $n = 36$, $p = 0.80$, Figure 4).

DISCUSSION

Our study explored how short-term low levels of artificial light at night may affect immediate responses in plant photobiology and herbivore interactions. It is important to note that this study represents a brief novel environment akin to new lights being installed in an environment, and not long-term exposure. Contrary to our predictions, grass grown under low-level urban light conditions after 19 days did not have higher net photosynthetic rates than those grown under starlight, nor did stomatal conductance, dark respiration, or ΦPSII differ significantly between light treatments. However, plants under urban light conditions grew taller than plants grown under starlight conditions during the initial 21 days of growth before crickets were introduced. Additionally, we found no evidence that crickets under urban light consumed more plant matter than crickets in starlight treatments, and survival rates of crickets did not differ between treatments. The results from this study suggest that short-term exposure to low levels of ALAN may not have significant effects on grass photobiology but may affect plant height.

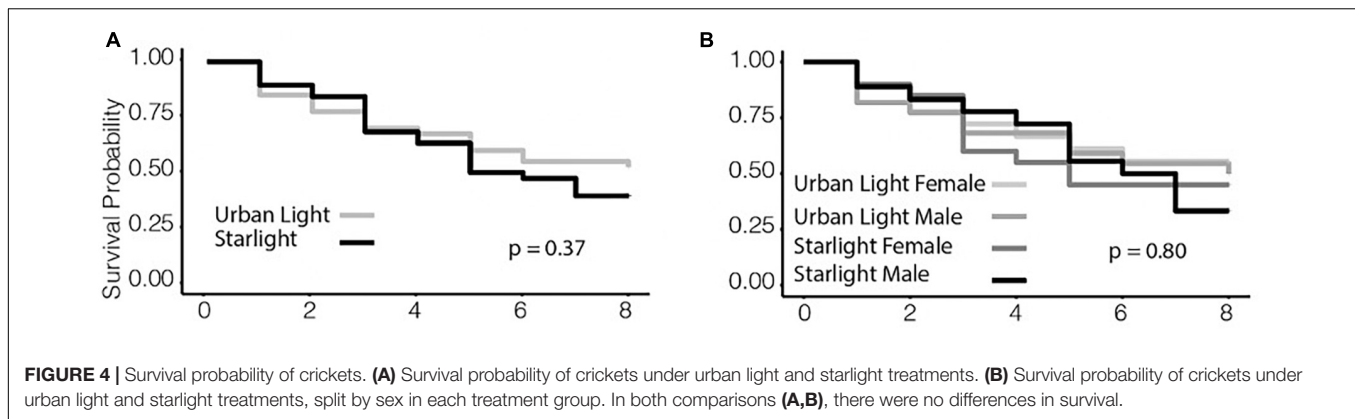
Studies investigating grass responses to higher levels but similar durations of ALAN illumination (e.g., $4 \pm 1 \mu\text{mol m}^{-2} \text{ s}^{-1}$ or 296 lux) found that plant photoreceptors were sensitive to small fluxes in light levels, which can change flowering phenology (Thimijan and Heins, 1983; Shin et al., 2010; Bennie et al., 2016). Many flowering plants require dark photoperiod signals to induce flowering (Bennie et al., 2016) such that light pulses, even one minute long, are enough to change their phenology (Parker et al., 1952; Singhal et al., 2018). The lower levels of light tested here were likely not bright enough to induce these changes in bluegrass and may have allowed bluegrass to properly detect photoperiod. Furthermore, plants often use nighttime darkness to repair damage from UV rays, suggesting the low levels of ALAN in our treatments may be dark enough for plants to continue to repair damaged cells and photoreceptors (Singhal et al., 2018). Moreover, net photosynthesis is a dynamic measurement that can vary within samples due to time and day (Miller et al., 1996) and our single measurement at the end of week 3 may not have captured treatment differences occurring at other times.

We found no difference in stomatal conductance or respiration between plants grown in urban light and starlight. Other studies of similar 4–6 week duration did note differences in stomatal density and stomatal opening and closing in the presence of ALAN at levels from $0.1 \mu\text{mol m}^{-2} \text{ s}^{-1}$ to $1 \mu\text{mol m}^{-2} \text{ s}^{-1}$ of blue and red light (Takemiya et al., 2005; Shimazaki et al., 2007). Another study found that yellow-poplar trees



exposed to ALAN (high pressure sodium lighting ranging from 82 lux to 4100 lux) for three years resulted in reduced nighttime stomatal conductance (Kwak et al., 2018). Given that we did not find any effects of ALAN on plant gas exchange, it is possible

that our light levels were too low, or grass was not subjected to our light levels for a long enough duration, to induce such responses. Reduced chlorophyll and rubisco concentration has been observed in phytoplankton grown under low light levels (6.6



lux; Poulin et al., 2014), and light as low as 3.5 lux has induced flowering in tree species across the United Kingdom (Ffrench-Constant et al., 2016). We also observed no treatment effects on photosystem II efficiency despite other studies noting adverse reactions in these physiological responses to light pollution (Zhang and Reisner, 2019; Meravi and Prajapati, 2020). Kentucky Bluegrass might be more adaptable to changing light regimes given that it is commonly used as a turf grass selected for its resilience to drought and heat stress (Wang and Huang, 2004). We observed a faster growth rate for grasses grown under urban light conditions compared to starlight conditions before the introduction of an herbivore. Plant growth rate is determined by a variety of factors, including, but not limited to, photosynthetic rate, specific leaf area, leaf lifespan, leaf mass fraction, and nitrogen absorption rate (Campbell and Grime, 1989; Poorter et al., 1991; Osone et al., 2008). Although we found no difference in net photosynthetic rate between treatments, growth rate differences could have been due to greater allocation to leaf area in urban light (Poorter and Remkes, 1990) although we did not measure such attributes. Further, our ALAN levels of 0.3 lux, though extremely bright, still fall within the range of the natural lunar cycle, occurring during rare, very clear nights with full moons (Gaston et al., 2013); thus, bluegrass may have been well suited to handle the ALAN levels tested.

ALAN is known to alter photoperiod detection in multiple organisms (Bennie et al., 2016), and these changes in photoperiod can impact plant growth and flowering (Cathey and Campbell, 1975; Blanchard and Runkle, 2010; Basler and Körner, 2012; Craig and Runkle, 2016). Increased growth and biomass have been noted in *Poaceae* species when exposed to high levels of ALAN ranging from 0.349 to 1.145 μmol

$\text{m}^2 \text{sec}^{-1}$ from metal halide bulbs (Flowers and Gibson, 2018), which corresponds to approximately 24.78–81.30 lux (Thimijan and Heins, 1983). However, after introduction of the herbivore, we observed no physiological responses in Kentucky Bluegrass, including no change in biomass. Photoperiod detection may not have been disrupted at our lower levels of ALAN, or it may have caused undetectable or non-measured physiological responses.

While animals rely on plants as a food source and shelter, we found no evidence that short term, low-level light pollution would impact these typical interactions between plants and insects. Artificial light at the level of 0.3 lux was not enough to induce changes in the amount of plant matter consumed by crickets or their survival, but light pollution at higher levels for longer periods of time could modify these interactions

TABLE 2 | ANOVA table comparing mean grass height at day 21 across light treatments and blocks.

	Sum of squares	df	Mean square	F	p
Light treatment	3.50	1	3.50	5.63	0.021*
Block	7.87	6	1.31	2.11	0.064
Residuals	39.8	64	0.622		

*Indicates a significant response (Light treatment, urban light $n = 36$, starlight $n = 36$).

TABLE 3 | ANOVA table showing the effects of light treatment, cricket treatment, and block (plus interactions between light and cricket treatment and cricket and block treatment) on daily percent change in grass height between day 21 and the end of the experiment.

	Sum of squares	df	Mean square	F	p
Light treatment	0.14	1	0.14	1.60	0.21
Cricket treatment	2.82	1	2.82	32.04	5.3×10^{-7} *
Block	0.85	6	0.14	1.62	0.16
Light: Cricket	0.002	1	0.002	0.023	0.88
Cricket: Block	0.90	6	0.15	1.70	0.14
Residuals	4.93	56	0.088		

*Indicates a significant response (Light treatment, urban light $n = 36$, starlight $n = 36$; Cricket treatment, present $n = 36$, absent, $n = 36$).

TABLE 1 | MANOVA table of the gas exchange results evaluating differences in photosynthesis, stomatal conductance in dark, stomatal conductance in light, fluorescence, and photosystem II efficiency (Urban light, $n = 11$; starlight, $n = 11$).

	df	Pillai	f	p
Treatment	1	0.18	0.45	0.83
Block	3	0.95	1.09	0.40
Residuals	17			

(Gaston et al., 2013; Macgregor et al., 2015; Bennie et al., 2016; Knop et al., 2017).

Overall, our research detected few changes to plant physiology under short-term exposure to low levels of urban light, suggesting that low levels of ALAN may not be as harmful to community interactions as predicted, at least initially. With rapid increase in human development, new lights are being introduced to unlit environments. Our experimental conditions may be representative of environments recently exposed to ALAN, such as a new housing development or newly urbanized areas. Other studies conducted at high levels of ALAN suggest artificial light can induce large changes in physiology and community interactions (Longcore and Rich, 2004; Gaston et al., 2013; Seymoure et al., 2019a). There may be a threshold level and length of exposure at which artificial light becomes harmful, causing detrimental effects to individual and ecosystem function with additional increases in intensity and duration. Understanding and identifying this threshold would allow for more effective management of night skies and natural light conditions (Dominoni et al., 2020). With estimates suggesting two thirds of Key Biodiversity Areas experience ALAN (Seymoure et al., 2019a; Garrett et al., 2020), it is important to identify the level at which artificial light becomes harmful and how natural night skies can be managed.

DATA AVAILABILITY STATEMENT

The raw data supporting the conclusions of this article will be made available by the authors, without undue reservation.

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AUTHOR CONTRIBUTIONS

MC, CB, NL, and BS conceived and designed the experiment. MC, CB, LA, and BS received funding for the study. MC and CB ran the study and collected data. MC, CB, and RG-N measured plant physiology. MC, CB, RG-N, NL, and BS analyzed the data. MC and CB wrote the manuscript with revisions from RG-N, LA, NL, and BS. All authors contributed to the article and approved the submitted version.

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